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A Survey of Pollution on Selected Streams in the Black Hills of South Dakota

Thomas J. Jurgens

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A SURVEY OF POLLUTION ON SELECTED

STREAMS IN THE BLACK HILLS

OF SOUTH DAKOTA

BY

THOMAS J. JURGENS

A thesis submitted in partial fulfillment of the requirements for the degree Master of Science, Major in Wildlife Biology, South Dakota State University

1968

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A SURVEY OF POLLUTION ON SELECTED STREAMS IN THE BLACK HILLS OF SOUTH DAKOTA

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A SURVEY OF POLLUTION ON SELECTED STREAMS IN THE BLACK HILLS OF SOUTH DAKOTA Abstract

THOMAS J. JURGENS

Under-the supervision of Dr. John Nickum

Seven streams in the Black Hills of South Dakota were surveyed to determine the influence of suspected sources of pollution on these streams.

The sources of pollution included both sewage treatment plant effluents and mining wastes. A comparison of the benthic fauna community below a pollution source to that above it was the primary basis for evaluating the effect of the pollution source on the stream.

The results of the benthic fauna samples indicated that the streams surveyed were being polluted. The degree of pollution of each stream was also indicated by these results. Chemical analysis were used to verify the results of the benthic fauna samples. These analyses concurred with the benthic fauna results and indicated the streams were being polluted.

ACKNOWLEDGMENTS

The author wishes to express his sincere appreciation to the many individuals who contributed to this study.

I want to thank R. Keith Stewart for his advice and counsel. His knowledge and experience of the Black Hills, which he generously shared, were helpful in initiating and conducting the study.

A sincere thank you goes to my adviser, Dr. John Nickum, for his cooperation and guidance offered in preparation of this thesis. I also want to thank Dr. Norman Schoenthal, formerly of the Wildlife Department, for his suggestions and ass�stance during. his tenure at South Dakota State University.

The microphotographs of representative macroinvertebrates shown in Figures VII, VIII, IX, and X were taken by Roger Woo, of the University of Minnesota, Limnological Research Center.

I especially wish to thank my wife, who typed this thesis, and whose encouragement during the study and assistance with the preparation of the manuscript were sincerely appreciated.

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INTRODUCTION

Gold mining was the first major cause of widespread pollution in the Black Hills. More recently .activities associated with mining, construction, waste disposal, and land and forest utilization (Figures I and II) have resulted in water pollution. A loss of over 1,000 miles of streams from the trout fishery in the last twenty years may be attributed to pollution (Stewart, 1961).

The major pollution problem currently degrading streams appears to be organic pollution. The sources of this pollution are stream-side . . homes and municipalities with insufficient sewage treatment facilities. Many homes adjacent to streams have only crude facilities for waste treatment. The wastes reach the stream either by direct deposition or indirect leaching. Community growth in the Black Hills area has resulted in the overloading of municipal sewage treatment facilities. This problem is compounded by an influx of tourists during the summer. When sewage treatment plants become overloaded, operators are forced to either partially treat wastes or allow raw sewage to by-pass the plant. These partially treated or raw wastes contain organic matter and toxic substances which reduce water quality (Figure I).

Consolidation of gold mining operations in recent years has limited pollution from this source to one drainage. However, potential mining pollution problems have been created in other drainages by reopening old gold mines with the expectation of discovering new minerals. Bog iron mining, recently made profitable by new advances in mining and new uses of this ore, has created a new pollution threat.

These mining operations are located adjacent to streams, where careless exploitations of their minerals could result in the destruction of several miles of streams.

Construction of roads and homes also has augmented the demise of streams. Roads designed to follow streams can be built at lower costs than those involving construction through mountainous terrain. Construction and maintenance of stream-side roads result in the introduction of large amounts of silt. This type of construction often necessitates direct modification of stream channels, such as rechanneling and straightening, resulting in a loss of stream length. Many of these modifications also result in accelerated erosion because flow rates of water are increased and vegetative cover that stabilizes stream banks is destroyed (Figure II). Construction of homes, primarily excavation and landscaping, also adds silt into the stream as excess soils are usually deposited in or adjacent to streams to avoid removal expenditures (Figure II).

Pollution from all these sources is intensified by reduced stream flows because pollutants are not adequately diluted. Orr (1959) reported a trend towards reduced stream flow caused either by dog-hair stands of ponderosa pine (Pinus ponderosa) or changes in precipitation patterns. Moisture is retained in the branches of dog-hair timber, where it evaporates and is prevented from reaching the ground; consequently, this moisture cannot reach the stream (Figure III). Drouth conditions can also result in reduced stream flows and intensify pollution because of the lack of dilution. Further evidence of reduced

stream flows is recorded in the files of Cleghorn Springs Trout Hatchery located on Rapid Creek. These records show a reduction in flow from nine million gallons per day in 1928 to four million. gallons per day in 1964.

Although pollution is generally apparent in the Black Hills, studies concerning the problem have been limited. The South Dakota Department of Health has reported pollution findings on Whitewood Creek (Anonymous, 1959); the Belle Fourche River {Anonymous, 1960); and Rapid Creek (Anonymous, 1964). The primary information reported in these studies concerns environmental health, and specific information regarding bottom organisms is briefly summarized or appended to chemical data. Other studies dealing with pollution have been reported by Stewart and Thilenius (1964) and Thilenius (1965).

The objectives of this study were: (1) to survey suspected sources of organic and mining pollution on major Black Hills' streams; (2) to determine the effect of these suspected sources on the streams by using benthic organisms as the main indicator of stream conditions; (3) to determine the practicality of using macroinvertebrates as a method of determining and monitoring stream conditions in the Black Hills.

The importance of macroinvertebrates as a tool in pollution investigation was emphasized by Hynes (1965) when he stated that a very simple study of the invertebrates can be used to determine the extent of pollution. Hynes (196Q) also pointed out that some of the advantages of using macroinvertebrates in studying pollution are: (1) a single series of samples reveals the state of animal communities (2) animal communities provide a more or less static record (3) biological records

show the result of intermittent pollution. It should be pointed out that macroinvertebrates are considered just one tool for pollution investigation, with best results obtained by using both biological and chemical methods.

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Upper photo showing cattle grazing on streambanks.
photo showing effluent release from Rapid City Figure I. Lower Sewage Treatment Plant.

Figure III. Top photo showing a stand of dog hair timber with a snow depth of 1.5 feet. Lower photo showing open area with a snow depth of 3.0 feet.

THE STUDY AREA

The Black Hills is a mountainous area lying along the South Dakota-Wyoming border. It encompasses an area of approximately 20, 600 square miles of which 12, 700 square miles-are in South Dakota. The area is drained by a large number of relatively small streams (Black Hills Area Resources Study, Anonymous, 1967). In the South Dakota portion of the area streams radiate from the main divide, which is along the crest of the limestone plateau that is generally adjacent and parallel to the South Dakota-Wyoming border (Newport, 1956). Figure IV graphically represents the geologic formations of the South Dakota portion of the Black Hills and also the location of sampling sites.

The following major streams in the Black Hills were surveyed: Spearfish, Rapid, Castle, Spring, Battle, and French Creeks, and Fall River. The geology of the region influences the physical, chemical, and biotic characteristics of these streams. The central portion of the Black Hills is composed of granite, and is surrounded by concentric rings of slates, limestones, and sandstones. Streams originating in limestone formations are more productive than those originating in granitic or slaty outcrops. All streams sampled originate in limestone except French and Battle Creeks. Only Rapid Creek and Fall River flow continuously to the Cheyenne River, while others studied become subterranean when they reach the eastern limestone rim.

Rapid Creek has the largest area of any drainage system in the Black Hills, and an average stream flow of 30. 9 cubic feet per second (cfs). (Detailed information regarding stream flows is presented in

GEOLOGIC AND STATION LOCATION MAP \equiv

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Table 1.) Rapid City uses this stream as a water supply source and also for waste removal from the municipal sewage treatment plant. (Detailed information regarding sewage treatment plants is presented in Table 2.) The sewage treatment plant has a capacity of 4×10^{6} gallons per day. When the load exceeds this maximum, partially treated and raw sewage are allowed to by-pass the plant. Three sampling stations were established on lower Rapid Creek--one above and one below the sewage treatment plant, and one ten miles downstream. The downstream station was established to measure stream recovery. Small amounts of vegetation, mostly periphyton were present in the upper and lower stations, Large deposits of organic sludge were common in eddy waters below the sewage treatment plant, but fast-flowing water kept riffle areas relatively free from sludge accumulations.

Spearfish Creek is considered **by** �iologists and many fishermen as the best stream in the Black Hills, having an average stream flow of 42.3 cfs. It flows throughout its entire course over limestone formations, with surface flow being maintained **by** a series of diversion dams and piping. Stream water is used by the town of Spearfish for potable water and to remove effluent from the Spearfish sewage treatment plant. One station was established above and one below the effluent outfall. The bottom at both stations was composed primarily of rubble with small amounts of sand and silt.

Spring Creek flows into.Sheridan Lake, one of the most popular ' recreation areas in the Black Hills. Average stream flow is 3.7 cfs, This stream receives wastes from the sewage treatment plant in Hill

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City. Two stations, one above and one below the sewage treatment plant, were established in the stream. Rubble was the predominant bottom material at both stations, with silt and aquatic vegetation present only at the lower stations.

French Creek flows only a short distance from its source before it flows through the town of Custer. During dry seasons the stream is intermittent above the town and the effluent from the Custer sewage treatment plant comprises the entire stream flow. Four miles downstream from Custer the stream enters Stockade Lake, which acts as a stabilization pond for any untreated wastes. Water released from Stockade Lake continues flowing until it reaches an area known locally as "the narrows". At this point it goes underground, but later resumes a surface flow for a short distance before it again becomes subterranean.

Five stations were established on French Creek to determine the modifying influences of an impoundment and underground flow on stream recovery following organic pollution. Sampling stations were located as follows: above Custer, below Custer, below Stockade Lake, above "the narrows" and below "the narrows". Bottom types were composed of rubble above and below the sewage treatment plant with sand at the other stations. Small amounts of aquatic vegetation were present above and below the sewage treatment plant and abundant below Stockade Lake.

Fall River is located in the southern part of the Black Hills. This stream originates in warm springs and has an average stream flow of 27.1 cfs. The streambed is composed entirely of limestone formations. The town of Hot Springs adds effluent from its sewage treatment

plant, One station was established above and one below the effluent outfall. The bottom at both stations is comprised primarily of sand which has been slightly solidified by calcarious deposits and a small amount of silt was also present.

The possible influence of bog iron mining on macroinvertebrates was investigated on the south fork of Rapid Creek. Two deposits of bog iron have been mined--one is adjacent to the south fork, and the other is on Hop Creek, a small tributary to the south fork (Figure V). Five stations were established in the mining area, including one above and one below both mining areas which are located approximately one-half mile from the confluence of the south fork with Hop Creek, and one station was established one-quarter mile below the confluence. The bottom type of the south fork is rubble and sand with no aquatic vegetation. The bottom type in Hop Creek was sand and silt at the sampling stations, but bedrock constitutes the bottom in the mined area.

Castle Creek is a primary tributary to Rapid Creek. It flows through extensive areas of unmined bog iron deposits. Three stations were established in Castle Creek to check the possible influence of these unmined deposits on macroinvertebrates. Stations were located t above, in, and below the main bog iron deposits.

Battle Creek is a small stream located in an abandoned gold field. Recently one of the old mines was reopened to mine beryllium, from which mine tailings are being deposited adjacent to the stream (Figure V). Stations were located above and below the mine, The bottom of both stations is almost entirely sand with no aquatic vegetation at either station.

Upper photo showing Hop Creek mining area. Lower photo show-
ing the beryllium mining area on Battle Creek. Figure V.

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Table 1. Population, Sewage Treatment Plant Capacity, Type Treatment, and Flows Through Sewage Treatment Plants of Streams Involved in the Study

All flows shown in gallons per day (g/d)

Indicates sampling months \mathbf{r}

Flows not actually recorded, but estimated by plant operators ri ri

Flows recorded by equipment it it it

	Fall River***		Battle Creek***		French Creek**		Spring Creek***
	1963	1964	1963	1964	1963 1964	1963	1964
Jan.	24.1	22.0	1.25	1.94		1.74	1.94
Feb.	23.8	22.5	2.11	3.03	1.9	1.90	3.03
March	25.6	23.4	7.11	2.57		2.05	2.57
April	23.6	24.5	19.4	5.82		3.47	5.82
May	21.3	26.1	22.9	12.5		12.9	12.5
June	23.6	24.0	85.5	29.1		94.3	29.1
July	22.5	22.6	21.3	25.7		29.7	25.7
Aug.	22.3	22.6	3.9	6.16	3.7	5.27	4.16
Sept.	27.6	23.8	6.1	1.81		2.93	1.81
Oct.	26.5	24.0	\cdot 2.93	1.25		2.93	2.80
Nov.	25.6	25.1	3.0	1.32		3.0	1.89
Dec.	24.5	22.9	2.24	1.32		2.24	1.52
Maximum							\bullet
discharge	74	44	300	131		171	33
Minimum							
discharge 16		18	0.8	0.4		0.5	0.8
Yean							
discharge 24.3		23.6	14.8	7.54		13.5	2.89
Annual Average							
discharge	27.1		\mathbf{x}		3.79		
Drainage area	137 sq. mi.		66 sq. mi.		199 sq. mi.		

Table 2. Monthly Average Maximum and Minimum Flows of Streams Involved in the Study

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		Castle Creek***		Hop Creek**		Spearfish Creek***		Rapid Creek in Rapid City ^{TT#}		
	1963	1964	1963	1964	1963	1964	1963	1964		
Jan.	2.02	2.34			25.9	39.5	16.3	29.8		
Feb.	2.11	2.32		1.0	30.4	38.4	19.0	27.6		
March	2.26	2.17			36.8	37.6	21.2	31.1		
April	2.0	13.2			109	60.7	27.6	65.6		
Мау	\cdot 2.17	21.2			105	88.6	40.1	107.0		
June	9.17	26.5			111	172.0	106	190.0		
July	8.29	18.4			57.5	69.9	121	115		
Aug.	7.66	22.2	2.3		40.3	59.1	55.1	73.8		
Sept.	7.65	23.9			41.8	50.4	54.8	46.0		
Oct.	2.37	12.8			38.2	50.3	39.0	49.6		
Nov.	2.38	2.20			35.4	49.6	30.8	33.8		
Dec.	2.18	2.47			36.7	49.0	29.6	32.5		
Maximum										
discharge	14	64			438	1,480	180	250		
Minimum										
discharge	1.9	2.0			20	31.0	12	9.4		
Mean		\bullet .								
discharge	4.19	12.5			55.7	63.7 [°]	46.8 [°]	66.8		
area	96 sq. mi.			168 sq. mi . 410 sq. mi.						
Annual average discharge Drainage	8.68				$\mathcal{A}^{\mathcal{A}}$ and $\mathcal{A}^{\mathcal{A}}$ and	43:3		60:3		

Table 2. (continued)

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All flow values shown in cubic feet per second (c.f.s.)

* Information unavailable
** Information compiled fro

** Information compiled from Surface Water Records of North and South Dakota, 1962, 1963, 1964
*** Records not available: flow determined at the time of sampling only Records not available; flow determined at the time of sampling only

NETHODS AND HATERIALS

Twenty-four sampling stations were established in the study area. Bottom samples were collected from riffle areas with a square foot Surber bottom sampler. The Bioassay and Pollution Ecology, Training Course Manual (Anonymous, publishing date unknown), states: (a) the riffle is one of the most satisfactory habitats for comparing stream conditions at different points; (b) the well-known square foot Surber sampler is one of the best quantative collecting devices from riffle areas; (c) at least two or three square foot samples should be taken at each station to insure that a reasonable percentage of the species present will be sampled. An attempt *to* reduce variation was made by selecting sampling sites with as many similar characteristics as possible. Cordone and Kelley (1961) list depth, velocity and substrate type as the significant features when considering sampling sites. Gaufin, Harris and Walter (1956) suggest that bottom forms are not randomly distributed and that bottom types to be sampled must be carefully selected if a small number of samples are to present a comprehensive picture of the fauna.

Two series of samples were collected for the study. One series of samples was collected during August, 1963 {summer samples). The summer samples consisted of one Surber sample collected from each site. Another series of samples was collected during February, 1964 (winter samples). Two Surber samples were collected on consecutive days at each station during the winter period.

After collection, organisms were sorted from debris by using a U. S. Standard Sieve Series, and preserved in a formalin solution. Final processing included separation, identification, and enumeration of individual organisms.

References used for identification included Review of Ephemeridae (Epherneroptera) in the �lissouri River Watershed with a Key to Species (Hamilton, 1959), Fresh-water Biology (Edmondson, 1959), Larvae of Insects, an Introduction to Nearctic Species (Peterson, 1960), and Aquatic Insects of California with Keys to North American Genera and California Species (Usinger, 1963). Nomenclature of organisms is according to Fresh-water Biology (Edmondson, 1959). No attempt was made to identify any adult forms such as Coleoptera and Hydracrina collected incidentally with bottom organisms.

Pollution evaluation by means of macroinvertebrates is simplified by establishing groups of organisms that react with some degree of similarity when affected by pollution. Three categories--pollution sensitive, intermediate, and tolerant--were established to evaluate this study. Organisms were classified on the basis of other studies, including Thelenius (1965), South Dakota Department of Health on Rapid Creek (Anonymous, 1964), and Brinkhurst (1963). These studies were used as a basis of comparison because they involved sources of pollution similar to those being investigated in this study. Studies on the environmental �equirements of Plecoptera (Gaufin, 1965); Ephemeroptera (Leonard, 1965); Tricoptera (Robak, 1965); midges (Curry, 1965); and Tubificidae (Brinkhurst, 1965), were also considered in classifying

organisms. These studies described the effects of factors such as dissolved oxygen, siltation, current, etc., on macroinvertebrates under both field and laboratory conditions.

The similarity between samples was determined by using Sorensen's coefficient of similarity

$$
K = \frac{2w}{a + b}
$$

where w equals the total of the smaller number of individual organisms taken at both stations; a equals the total number of organisms at the first station; and \underline{b} equals the total number of organisms at the second station (Phillips, 1959). Samples having completely different numbers and kinds of organisms would have a similarity index of zero; samples which were identical in both numbers and kinds of organisms would have a similarity index of 100.

Indices of similarity were determined between samples taken above and below suspected pollution sources for both summer and winter samples. Winter samples taken from the same relative location on consecutive days were also analyzed to determine similarity indices.

Chemical data, presented in the results section, was collected in association with other stream studies in the Black Hills area. This data is presented only from samples which were taken from stations that closely coincided with bottom sampling stations; therefore, data is lacking for some stations.

Water samples were analyzed by Inland Analytical Laboratories, Inc., in Rapid City, South Dakota, using methods described in Standard Nethods for the Examination of Water and Wastewater for the following:

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Figure VI shows typical winter and summer sampling sites. Figures VII, VIII, IX, and X are microphotographs of some of the representative organisms that were sampled. \sim \sim

Figure VI. Upper photo showing typical summer sampling site. Lower
photo showing typical winter sampling site.

Figure VII. Microphotographs of Tricoptera. Top photo showing
Hydropsyche. Lower photo showing Glossosoma enclosed in a case.

Figure VIII. Microphotographs of Diptera and Crustacea. Top photo showing Simulidae larva. Middle photo showing two forms of tendipeds. Lower photo showing the Crustacea Hyallela.

Figure IX. Microphotographs showing dorsal view (upper photo) and ventral view (lower photo) of the Mayfly nymph; Ameletus.

Figure X. Microphotographs showing dorsal view (upper .photo) and ventral view (lower photo) of the Stonefly nymph, Acroneuria.

RESULTS

Summer samples collected from Rapid Creek above the Rapid City sewage treatment plant contained three sensitive genera: Tricorythodes spp. Ulmer, Centroptilum spp. Eaton, and Baetis spp. Leach. One intermediate form Lumbricidae also appeared above the plant. These organisms were all eliminated below the plant. Tolerant organisms including Glossiphonia spp. Johnson, Limnodrilus spp. Claparede, Psychoda sp. Latreille, and tendipeds (Family Tendipedidae--midge larvae) with anal gills were present below the plant. Tendipeds were divided into two groups--those with gills on the tenth abdominal segment, and those without such gills. According to Stewart (1965), these abdominal gills enable tendipeds to withstand much lower dissolved oxygen concentrations typical of polluted areas. Tendipeds with abdominal gills were classified as pollution tolerant, while those without were considered intermediate. At the station established ten miles downstream, sensitive genera of Tricorythodes spp. and Paraleptophlebia spp. Lestage returned and Neocloeon spp. Traver and Hydropsyche spp. Pictet appeared. Intermediate forms present at the downstream station were Hyallela sp. Saussure and tendipeds without anal gills.

Winter samples were similar to those collected during the summer; sensitive organisms sampled above the plant were Tricorythodes spp., Bae tis spp. , and Hydropsyche spp. These organisms were absent below the plant. Intermediate forms including Simuliidae, Hyallela sp., and tendipeds without anal gills were found above the plant. Intermediate forms present below the plant were Eclipidrilus sp. Eisen and

tendipeds without anal gills. At the downstream station Baetis spp. and Hydropsyche spp. returned and Cheumatopsyche spp. Wallengren was present . Hyallela sp., an intermediate form, also returned. Tolerant genera were found at all three stations. Above the plant, Tubifex sp. Lamarck and Glossiphonia spp. were found, while below larger numbers of these genera and Helobdella sp. Blanchard, Psychoda sp., and tendipeds with anal gills were present. Tendipeds with anal gills and Helobdella sp. disappeared downstream and the numbers of Tubifex spp. and Psychoda sp. decreased. Additional benthic fauna data from the Biological Survey Report on the Rapid Creek Water Pollution Inyestigation is presented in Appendix A.

The coefficient of similarity indices for the stations above and below the sewage treatment plant are summer 0, winter first day 5, winter second day 5. Indices between the station above the plant and the downstream station are summer 16, winter first day 35, and winter second day 51. Index values on Rapid Creek for the consecutive days with comparable sampling sites are 74 above the sewage treatment plant, 70 below it, and 58 at the downstream station.

Organisms collected from Fall River showed variation above and below the Hot Springs sewage treatment plant. Summer samples showed a reduction of sensitive organisms from four above the sewage treatment plant to two below the plant. One intermediate form was taken above and two were taken below the plant. Winter samples showed a greater variation between stations. Twelve sensitive organisms were sampled above the plant and only two were sampled below. Intermediate forms decreased from two above to one below the plant. No tolerant forms were taken in
any of the samples; their absence can be explained by the lack of bottom habitat suitable for these organisms.

Chemical samples collected from Fall River show an increase of total solids from 430 ppm to 940 ppm; chloride from 106 ppm to 171 ppm; sodium from 160 ppm to 300 ppm; total phosphates from . 18 ppm to .96 ppm; ammonia from .33 ppm to 1.02 ppm; nitrite from .02 ppm to . 10 ppm; and nitrate from . 06 ppm to . 28 ppm. Complete chemical analysis is shown in Table 8.

The coefficient of similarity indices for the stations above and below the sewage treatment plant are 13 for the summer samples, 4 for the first day and 5 for the second day winter samples. The index values for comparative location samples on Fall River are 88 above the plant and 64 below it,

French Creek samples above and below the Custer sewage treatment plant showed only a slight variation in types of organisms. Summer samples showed a decrease in sensitive organisms from six above the sewage treatment plant to two below it. Winter samples did not show this variation; only the numbers of tendipeds without anal gills showed a decrease below the plant. Numbers of sensitive organisms increased at stations below Stockade Lake and in "the narrows" area. Kinds and numbers of intermediate and tolerant species did not vary appreciably in the French Creek stations.

A comparative chemical sample was not available from the station above the sewage treatment plant, but other stations showed a general decrease of constituents at each station below the sewage treatment

plant. Selected chemical values for French Creek stations are shown in Table 3. Additional chemical data from French Creek is presented in Appendix B.

Table 3. Comparison of Selected Chemical Constituents of the

Index of similarity values of the French Creek winter samples with similar locations are above the sewage treatment plant 64, below the plant 44, below Stockade Lake 79, above "the narrows" 79, and below "the narrows" 70. Table 4 shows the index of similarity values for the French Creek stations compared to the station above the sewage treatment plant.

*Sewage Treatment Plant

The reaction of the benthic community in Spring Creek below the Hill City effluent outfall was generally one of increase in both numbers and kinds of organisms when compared to the station above the sewage treatment plant. Sensitive organisms increased from four above the sewage treatment plant to six below it. Winter samples showed an even greater increase of from nine above the plant to 11 below it. Numerical increases of other forms are exemplified by Hydropsyche spp. , which increased from 120 organisms above the plant to 1, 399 below it, and by Cheurnatopsyche spp., which increased from 97 above to 571 below. Intermediate and tolerant forms reacted to the Hill City effluent the same way as the sensitive organisms showing increases in kinds and number of organisms.

Results of chemical analysis also showed increase in most constituents below the plant. Total solids increased .from 102 ppm to 307 ppm; total phosphate remained the same ; ammonia increased from . 80 ppm to 2. 02 ppm.

Indices of similarity values comparing the station above the plant to the one below are summer sample 13, winter sample first day 28, winter sample second day 17. Values comparing the same sites on consecutive days are 74 for the station above the sewage treatment plant and 86 for the station below it.

Samples from Spearfish Creek in general were very similar to those from Spring Creek. Sensitive organisms again showed increases in kinds and numbers. Intermediate forms also showed slight increases in kinds and numbers while tolerant species were almost entirely lacking.

Simuliidae showed large increases in the summer sample, from 200 to 2, 306, and tendipeds with anal gills showed a similar increase in the · winter samples, 45 to 331, above and below the effluent outfall.

Chemical data concurs with biological data and does not show any large increases in chloride, sodium, nitrite, nitrates; phosphates did show a slight increase from . 10 ppm above the plant to . 66 ppm below it. Additional chemical data from Spearfish Creek is presented in Appendix C.

Indices of similarity values comparing the station above the plant to the one below it are summer 16, winter first day 24, and winter second day 16. The index of similarity value for samples taken above the plant on consecutive days is 63, while the value for samples taken below the plant is 91.

Complete biological results for stations associated with organic pollution are shown in Tables 5, 6, and 7. Table 8 shows the complete chemical analysis for the stations associated with organic pollution.

Bog iron mining operations in the south fork of Rapid Creek and Hop Creek areas were sampled both biologically and chemically. · Bottom samples above and below the mine on the south fork were similar. No macroinvertebrates were collected in the lower Hop Creek station during either sampling period. Organisms were reduced in kinds and numbers in the south fork below its confluence with Hop Creek.

Table 5. Organic Associated - Summer Samples

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Table 6. Organic Associated - Winter Samples - First Day

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PLECOPTERA	FALL RIVER ABOVE SEWAGE TREATMENT PLANT	SENAGE BELOW TREATMENT PLAIT FALL RIVER	TREATMENT PLANT FRENCH CREEX ABOVE SENAGE	TREATMENT PLANT CREEK BELOW FRETCH SEMAGE	CREEK BELOW STOCKADE LAKE FRENCH	FRENCH CREEK ABOVE SYORROWS	FRENCH CREEK BELOW EFORE	TREATMENT PLANT CRIZEK ABOVE	SEWACE TREATMENT PLANT SPRING CREEK BELOW	SERNOE TREATMENT PLANT ULTID CREEK ABOVE	SEWGE TREATMENT PLANT RAPID CREEK BELOW	DOWNSTREAM STATION PAPID CREEK	SEWAGE TREATMENT PLANT SPENFISH CREEK ABOVE	SEHAGE TREATHENT PLANT SPEARFISH CREEK BELOW	
Acroneuria sp.														$\overline{\mathbf{z}}$	
Arcynopteryx spp. Isoperla spp. <u>Alloperla</u> spp. EPHEIEROPTERA Ameletus sp.	3					ı	8 28	53	70				27 $\mathbf{1}$		
Tricorythodes spp. Paraleptophlebia spp. Centroptilium spp. Neocloeon spp. Ephemerella spp.	з 2				9		14 5	6		3			18	195	
Baetis spp.	3					22	9		85	13		¹	57	6	
COLEOPTERA Narpus spp. Optioservus spp.	21						2	4	3 5					2	SENSITIVE
Zaitzevia spp. LEPIDOPTERA	4														
Elophila sp. TRICOPTERA	187	3							3						
Clossosoma spp. Chimarra spp. Agravles spp.	5							31	8						
Hesperophylax spp. Linnephilus spp. Leptocella spp. Oecetis spp.	4						4		4						
Triacnodes spp. Brachycentrus spp.			4	$\mathbf{1}$		4	ı	2	7				11	ı	
Helicopsyche spp.	15	11													
Hydropsyche Spp. Cheumatopsyche spp. AMPHIPODA	243 3		23			208 195	115 215	44 33	754 296	22		11 33	23 6		
Gammarus spp. Hyallela sp. ODONATA			70	68			ı	24		$\overline{\mathbf{c}}$		15			
Gomphus spp. Erpetogomphus spp. Ophiogomphus spp.							1		ı						
DIPTERA Simuliidae									з						
Tendipeds (with- out anal gills) Bezzia sp.	10		995	38	99	167	35	44	125	31	8	40	17	147	INTERNEDIATE
Chrysons spp. Tabanus sp							2	6	ı						
Tipula sp. Hexatoma sp. Atherix Sp.									ı					2	
PLESIOPORA Eclipidrilus sp.					ı	1	\bullet	5.	5.		15			12	
PLESIOPORA Limnodrilus spp.							20	2	4					11	
Tub <u>ifcx</u> spp. RHYNCHOEDELLIDA Helobdella sp.				2	\mathbf{v}			1			163	з			
Glossiphonia spp. DIPTERA												\mathbf{z}			
Tendipeds (with			ı	2					ı		84				
anal gills) Psychoda sp.											4				TOLERANT
SENSITIVE	12	2	2	2	2	5	10	7	10	Э	0	з	7	5	
INTERMEDIATE TOLERANT	2 ٥	٥ 0	2 ı	2 2	$\overline{\mathbf{z}}$ 1	2 ٥	5 ı	4 2	6 $\overline{\mathbf{2}}$	2 0	2 3	2 2	1 0	з 1	
TOTAL NAMBER OF	504	14	1993	110 127		598		467 255 1376		71	274 107		160	378	
CROADISHS															

Table 7. Organic Associated - Winter Samples - Second Day

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 0.07

255.0

0.09

320.0

0.46

290.0

1.12

580.0

 0.28

1200.0

 0.06

1200.0

NEG

285.0

NIL

425.0

 $19 - N03$

 $20 - S.C.$

NIL

410.0

NEG

230.0

Table 8. Chemical Results from Organic-associated Stations

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Chemical data from the three s'tations on the south fork of Rapid Creek show an increase in total solids from 250 ppm at the station above both mines to 303 ppm above Hop Creek and 250 ppm below Hop Creek; pH 7.75 above both mines, 7.8 above Hop Creek, and 7.1 below Hop Creek; sulfates 14 ppm above both mines, 45 ppm above Hop Creek, 63 ppm below Hop Creek; total iron was negligible above both mines, . 44 ppm above Hop Creek and . 73 ppm below Hop Creek. The station above the mine on Hop Creek did not have a comparable chemical station; however, chemical results below the mine showed 400 ppm total solids, pH of 3.2, 270 ppm sulfates and total iron of 11.2 ppm.

Indices of similarity values comparing the station above both mines to the one above the confluence with Hop Creek are summer 66, winter first day 21, and winter second day 37. Values comparing the station above both mines to the one below the confluence with Hop Creek are summer 0, winter first day 20, and winter second day 9. Index of similarity values for the three stations on the south fork of Rapid Creek are above both mine areas 33, above the confluence with Hop Creek 68, and below the confluence with Hop Creek 33.

Castle Creek was sampled in an unmined bog iron deposit area after trout mortality in the area was reported late in the fall. Samples were collected from Castle Creek only during the winter sampling period, Sensitive organisms did not show any appreciable differences at any of the three stations. Intermediate forms were reduced from 6 above the deposit area to O in the deposit area; three intermediate forms were present at the lower station. One tolerant

form was present in the upper station; none were present at the other two stations.

Chemical data shows increases in: total solids from 211 ppm to 252 ppm, turbidity 6 ppm to 42 ppm, sulfates 23 ppm to 78 ppm and total iron . 04 ppm to 4.5 ppm; from the upper station to the station in the bog iron area. The pH value at the lower station was 7.1 compared to 7.9 at the upper station. Values at the lower station of the constituents listed above returned to those of the upper station except for sulfates, which were 79 ppm.

Index of similarity values for the Castle Creek stations indicate population differences between stations. The values comparing the upper and mid-station are first day 9, second day 11, and those comparing the upper and lower stations are first day 15, second day 12. Similarity values on Castle Creek for the consecutive days with comparable sampling sites are above the bog iron deposits 46, in the deposit area 66, and below the deposit area 57.

Samples collected from Battle Creek showed a reduction in numbers and kinds of organisms below the beryllium mine. All genera of. Plecoptera, Ephemeroptera, and Coleoptera present above the mine were absent. " Numbers of all other organisms were reduced at the station below the mine. Additional benthic fauna data, collected by South Dakota Department of Game, Fish, and Parks personnel, is presented in Appendix D.

Chemical data from Castle Creek corresponds with the biological data and showed increases in many constituents. Increases from above

the mine to below the mine were recorded for the following constituents : total solids 170 ppm to 638 ppm; turbidity 4 ppm to 37 ppm; chloride 120 ppm to 237 ppm; sulfates 25 ppm to 225 ppm. The pH was lowered from 6. 3 above the mine to 3 .O below the mine.

Index of similarity values comparing the Battle Creek stations above and below the mine are summer 30, winter first day 31, and winter second day 10. The value comparing similar samples above the mine is 74 and the value comparing stations below the mine is 41.

Complete biological results for stations associated with mining areas are shown in Tables 9, 10, and 11. Table 12 shows the complete chemical analysis for the stations in the mining areas. Table 13 shows the index of similarity values for winter samples taken from the same relative area on consecutive days. Table 14 shows the index of values for stations above and below various suspected sources of pollution.

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Table 9. Mining Associated - Summer Samples

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Table 11. Mining Associated - Winter Samples - Second Day

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Table 12. Chemical Results from Mining-associated Stations

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Table 13. Index of Similarity Comparisons Between Winter Samples Taken From Same Relative Area on Consecutive Days

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Table 14. Index of Similarity Comparisons for Stations Above and Below

DISCUSSION

Hawkes (1964) discussed various aspects of pollution and macroinvertebrates, including how pollution affects the benthic community and the responses of the benthic community to pollution, which are . summari**z**ed by the following comments ,

Pollution can either affect the organism directly through some metabolic process or indirectly through habitat alteration. Several factors determine the influence of pollution upon the benthic community, including toxicity thresholds of organisms, reduction of food, elimination of predator species, and changes in composition of bottom materials. The riffle community is dependent on materials carried in by the current. Any changes in these materials will affect the community.

Macroinvertebrates react to organic pollution in one of the following *ways :* 1) Mild pollution result� in a general increase in most organisms, except for genera that are highly sensitive which will be eliminated. 2) Additional pollution will eliminate most organisms in the sensitive category, reduce the number of forms in the intermediate category, and those in the tolerant category will increase. 3) Severe pollution will result in the loss of organisms in the intermediate category, and an increase of organisms in the tolerant category.

Toxic and organic wastes usually exhibit similar effects on the benthic community, although certain species may be affected differently. Certain species show mome tolerance to toxic wastes, while others show

less tolerance; for example, some species of stoneflies are eliminated by a small amount of organic pollution, but can withstand large amounts of heavy metals, and certain species of Diptera have shown just the opposite reaction. However, when considering the entire benthic community, Hawkes concluded, the effects are very similar.

The "Report on Water Pollution Investigation Rapid Creek", December, 1963, page V, by the South Dakota Department of Health, clearly stated that Rapid Creek is polluted below the Rapid City sewage treatment plant:

"Clean stream water quality in Rapid Creek was found above Rapid City. The sanitary and industrial wastes at Rapid City are only partially treated. Repeated by-passing of raw munici-
pal wastes is contrary to health regulations. Improperly pal wastes is contrary to health regulations. treated waste water from municipal waste treatment facilities creates serious public health hazards and water-course degradation in the receiving stream. The physical, chemical, and biological quality of lower Rapid Creek waters precludes use of this water for safe beneficial purposes. "

Because Rapid Creek is knqwn to be polluted, it was used as a standard to determine the effect of pollution on macroinvertebrates and as a comparison for other streams sampled.

The sampling results generally agree with those published by the South Dakota Department of Health. The elimination of sensitive organisms and the occurrence of such species as Psychoda sp. below the sewage treatment plant indicates that Rapid Creek is being polluted by the effluent from the Rapid City sewage treatment plant. The occurrence of sensitive and intermediate organisms at the downstream station indicates that the stream is recovering from the heavy pollution immediately below the sewage treatment plant. The occurrence of these

organisms does not infer high quality water as the organisms present are the more resistant organisms.

The results of the French Creek samples indicate a change in water quality below the Custer sewage treatment plant. This minor . change in organisms probably does not reflect the full influence of the Custer plant because organisms above the plant are limited by low stream flows. The increase in kinds of organisms at each downstream station reflects the improvement of the water quality. However, water flows at these stations are more consistent because of releases from Stockade Lake and small feeder streams. This improvement may be the result of either distance from the sewage treatment plant as in the case of the downstream station on Rapid Creek or the influence of Stockade Lake .

The improvement in water quality at the station below "the narrows" from that above "the narrows" is probably due to the influence of the underground aquifer as the distance between the stations is approximately one-half mile and it is doubtful if distance alone could result in the improved water quality.

The results of chemical sampling verify those of the biological samples. The high values of total solids, turbidity, sulfates, phosphates, nitrites, and nitrates below the plant show that the Custer sewage treatment plant is adding to the pollution load of French Creek.

The degrading effects of the Custer effluent are not only apparent on French Creek, but also on Stockade Lake, one of the main

sources of water-based recreation to visitors at Custer State Park. This lake shows many signs of organic enrichment, or eutrophism, including heavy algal blooms, dense aquatic vegetation, an ooze bottom, and the inability to support a trout population as it once did. Mackenthun, Ingram, and Porges (1964) list one of the main methods of minimizing conditions leading to water enrichment as stopping the discharge of sewage and decomposable organic industrial wastes, which contain high concentrations of nitrogen and phosphorus, which will manifest in nuisance growths of aquatic plants.

The reduction in sensitive organisms below the Hot Springs sewage treatment plant shows that the water quality of Fall River is being lowered by the effluent from the plant.

Chemical samples collected from Fall River support the biological data. The increase in total solids, chloride, sodium, phosphates, nitrites, and nitrates correspon�s to the decrease in sensitive organisms.

Spring Creek samples indicate that the effect of the Hill City sewage treatment plant is one of enrichment of the stream. The presence of the stonefly larvae Isoperla spp. Banks indicates that Spring Creek is not being seriously degraded by the Hill City effluent.

Chemical samples did not show any major increases except for total solids and ammonia; and as the biological samples, they indicate enrichment of the stream.

The situation on Spring Creek is similar to that on French Creek in that the Creek flows into a major recreation reservoir,

Sheridan Lake. This reservoir is showing signs of eutrophication, especially in the inlet area where dense stands of aquatic vegetation are apparent.

Spearfish Creek samples were similar to those collected from Spring Creek in that the reaction was an increase in total number and kinds of organisms. This increase is indicative of the stream being enriched by the effluent from the Spearfish sewage treatment plant. The occurrence of the sensitive stonefly species Acroneuria sp. and Isoperla spp. below the sewage treatment plant is further evidence that the effluent is not causing serious degradation of the stream.

Chemical data showed slight increases in some constituents, indicating that the stream is being enriched by the effluent from the sewage treatment plant.

Samples from the bog iron mining area indicates that the mine adjacent to the south fork did not influence the water chemistry to cause any significant changes in the benthic fauna. Chemical samples did show an increase in iron; however, it did not cause the bottom organisms to change.

Samples taken in the Hop Creek area and in the south fork below Hop Creek did show major changes in both the biological and chemical samples. No organisms were taken below the mine in Hop Creek, iron was 11.2 ppm and the pH was lowered to 3.2 ppm.

Data from the station in the south fork below Hop Creek also showed that the Hop Creek mine was influencing the biological and chemical characteristics of the south fork. The elimination of most benthic organisms, the increase in sulfates and iron, and the lowering of the pH in the lower s tation in the south fork is evidence of the effects of the Hop Creek mine.

The effect of the high iron concentrations especially in feeder streams to reservoirs could result in a general decline in productivity of the reservoirs. Ruttner (1953) states when ferrous iron and phosphate occur together in the hypolimnion of a lake, an insoluable ferric phosphate is precipitated at times. There is some evidence that this phenomenon may be in effect in Pactola Reservoir which is fed by Rapid Creek.

Data from the Castle Creek stations show the effects of bog iron deposits, as did the south fork mining s tations. Although no mining has been done in Castle Creek, iron is leaching into the creek from deposits near the creek. Organisms decreased when the iron and sulfate content of the water increased in the iron deposit area. The organisms that appeared to be affected the most by the increased iron were those listed as intermediate. Many of the organisms that were eliminated did recur at the lower station corresponding to a decrease in iron and sulfate at the same station.

Battle Creek samples show the effect of the beryllium mine on the biological and chemical samples. Organisms were reduced in numbers and kinds at the downstream station. Chemical samples showed increases in almost every constituent and correspond with the reduction in benthic fauna to show the effects of the beryllium mine.

SUMMARY AND CONCLUSIONS

Results of macroinvertebrate sampling on each stream reflect changes in water quality; thus, each stream is being polluted by the suspected sources of pollution that were investigated. The reaction of the benthic community not only indicates that the streams are being polluted, but also the degree of pollution of each stream. Rapid Creek shows the greatest reduction in water quality due to organic pollution. Fall River and French Creek are also being severely polluted by organic wastes, although the main effects in French Creek are more serious on Stockade Lake than on the Creek itself. Spearfish Creek and Spring Creek are being only mildly polluted by sewage treatment plant effluents. The pollutants being added to Spring Creek are evidently accumulating in Sheridan Lake; thus, the mild pollution of Spring Creek must be considered as serious.

Mining is also responsible for degradation of streams. Hop Creek is grossly polluted by mining of bog iron, and this pollution is affecting the south fork of Rapid Creek. Castle Creek is being polluted by leaching from bog iron deposits; and if these deposits were to be mined similar to the Hop Creek area, the results could be the same as Hop Creek and the pollution extended further down stream. Beryllium mining and disposal of process wastes is polluting Battle Creek.

Chemical data also indicated that streams investigated are being pollu ted. This data concurs with and therefore supports the macroinvertebrate data. It is evident that the best pellution investigations

involve both biological and chemical evaluation; however, biologists are often limited by time, equipment, and budgets, and are unable to conduct thorough investigations. Macroinvertebrates are one tool that enables biologists to overcome some of the previously mentioned problems, and yet obtain valuable information regarding stream conditions. In many instances pollution investigations based on benthic communities can be simplified by limiting the identification of organisms to the order or family level. However, identification to the genus or species level is necessary in cases of mild pollution.

Macroinvertebrates can indicate the degree of pollution of a stream; however, they cannot indicate the chemical constituents causing pollution. In many instances the type of pollution is evident, such as sewage treatment plant effluents; however, in other instances, the composition of pollutants is unknown and can only be determined by chemical analysis. One of the most beneficial uses of macroinvertebrates would be as a monotoring device in streams; this would involve sampling of specific sites at regular intervals. Any significant changes in the benthic fauna could be an indication of a possible change in water quality, and would necessitate a more intensive investigation.

Records from this type of program would be invaluable for evaluating the effects of new pollution sources or evaluation of remedial measures applied to known pollution sources. For example, Rapid City is currently constructing a new sewage treatment plant, and the effectiveness of this plant could be determined by sampling macroinvertebrates before and after the start of its operation. Also, the

recovery of the stream below the old plant could be determined after it is no longer in service.

Detection and curtailing pollution is probably the main problem currently facing fisheries biologists in the Black Hills, Reduction of water quality by pollution has resulted in the loss of many miles of stream from the trout fishery, and impoundments now receive the majority of fishing pressure. Impoundments are also important for recreation, such as water skiing and swimming. These impoundments cannot continue to receive contaminants carried by their feeder streams and still maintain their high quality. This fact is evidenced by Stockade and Sheridan Lakes.

Construction of new dams in the Black Hills is limited in part by pollution. Attempts to select dam sites away from pollution often necessitates selection of sites high on the drainage where the water supply is insufficient or construction costs are prohibitive.

Continued lake pollution will affect the economy in the area of the Black Hills. The Black Hills are popular as a recreational area and also have many points of interest which attract tourists. For example, Mt. Rushmore and Custer State Park both average over one million visitors each year (Appendix E). Degradation of the lakes to the point where they are no longer attractive as a recreational source will decrease the ability of the area to retain people.

Pollution not only affects the recreational aspects of streams and lakes, but also the agricultural aspects. Water polluted by . organic or toxic wastes cannot ·be used effectively for irrigation or livestock.

Towns in the Black Hills use the streams as a water supply source and could be in danger of losing it if pollution continues. The water at least will require additional treatment, resulting in higher costs for potable water. Towns may be faced with not only low-quality water, but also with an insufficient supply, if the trend towards reduced stream flow is continued.

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APPENDIX

Appendix A. Benthic Fauna from Rapid Creek

Data taken from Biological Survey Report from stations above and below Rapid City sewage treatment plant.

Appendix B. Chemical Data from French Creek

Information obtained from Dept. of Game, Fish, and Parks files. All values in parts per million.

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Appendix C. Chemical Information Regarding Spearfish Sewage We will all Treatment Plant Effluent* المتواردة لاطاء فتتراط بمنتوب فللراز

Samples Taken from Spearfish Creek*

It can be seen that the existing facility is not meeting public health standards even under the optimum conditions of the test period. A more serious condition prevails during summertime peak loading when the receiving stream is down in flow and sewage flows at a maximum. $*$

*Information obtained from Preliminary Report, Waste Water Treatment Facilities for Spearfish, South Dakota.

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Appendix D. Benthic Fauna Data from Battle Creek*

*Data obtained from Game, Fish, and Parks Dept. files.

Appendix E. Attendance Figures for Leading Tourist Attractions in . th� Bl�ck Hills Ar�a .. \mathcal{L}_{max}

* Information obtained by personal correspondence with the Superintendent of Mt. Rushmore National Memorial.

** Information obtained by personal correspondence with the Superintendent of Custer State Park.

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*** Information obtained from Preliminary Report, Waste Water Treatment Facilities for Spearfish, South Dakota.